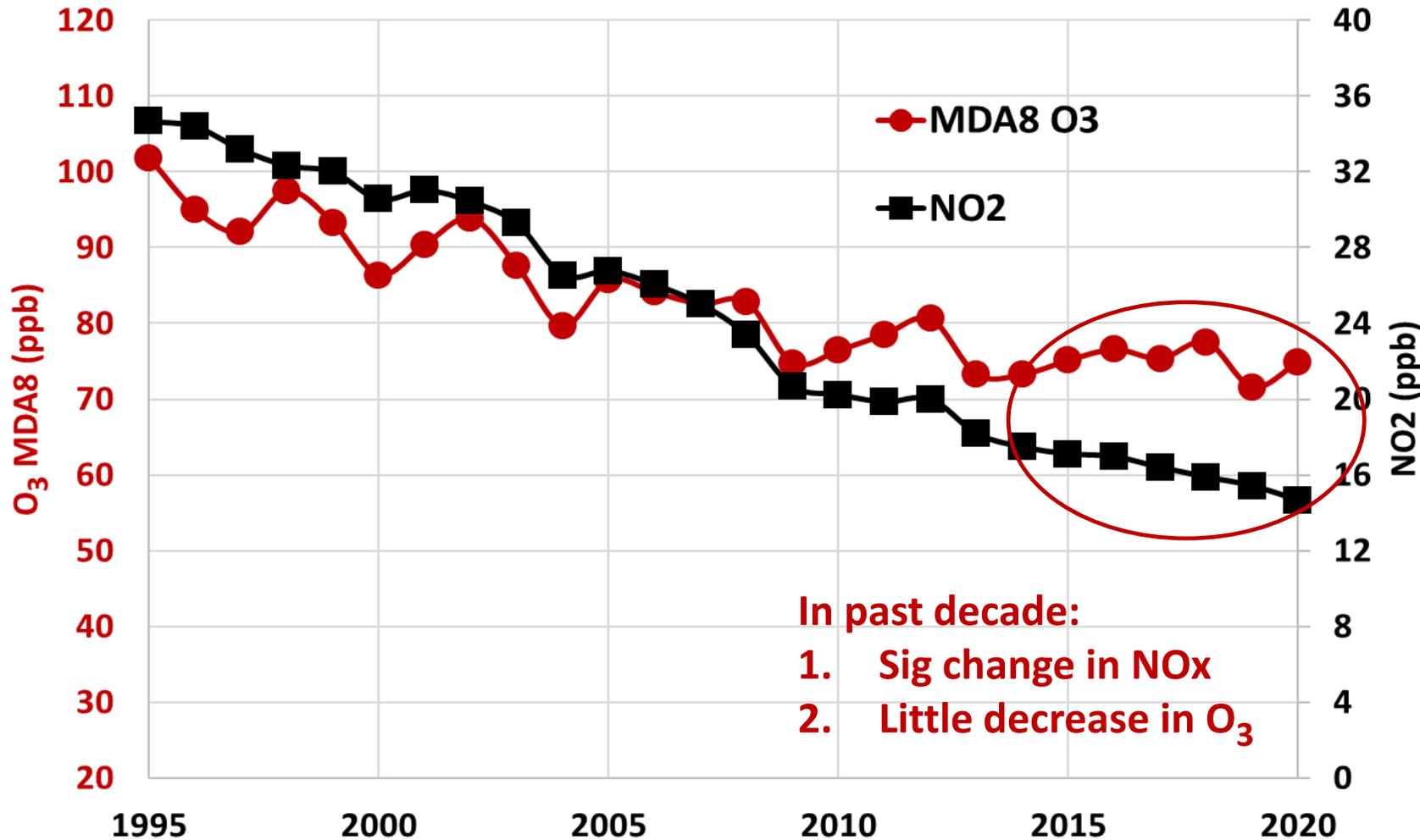


What can the pandemic tell us about NO_x and Ozone in the U.S.?

Dan Jaffe, University of Washington



MOTIVATION:

1. Given strong reductions in all of the primary pollutants over past decades why hasn't O₃ declined as much?

2. How much do we need to reduce NO_x to meet the O₃ standard?

3. Given 2020 COVID related reductions in VMT and NO_x, do we see the expected O₃ reductions?

Annual 4th highest MDA8 and May-Sept mean 1 hr daily max NO₂ for 32 U.S. Non-Attainment Areas (NAAs)

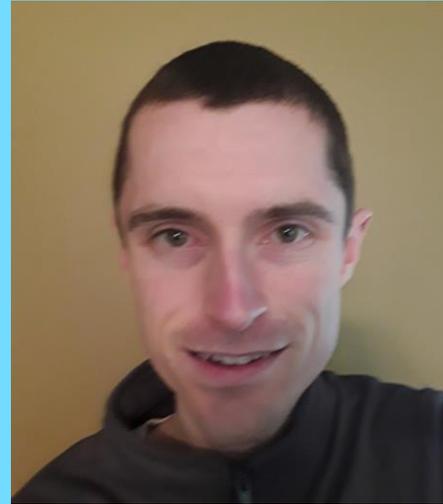


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former ug,
Mt. Bachelor obs
and analysis of low
cost sensor data**



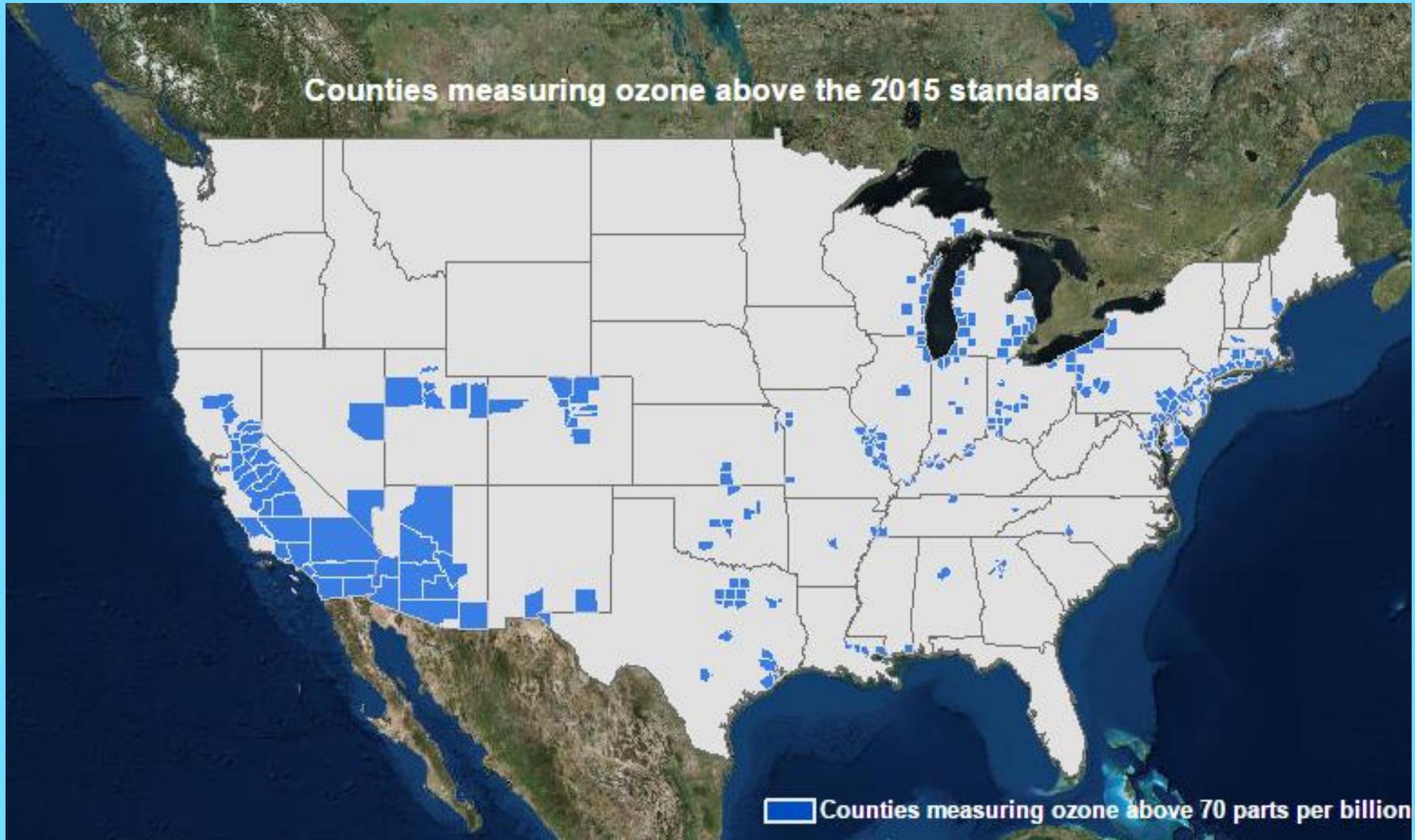
**Noah Bernays,
Mt. Bachelor site tech**

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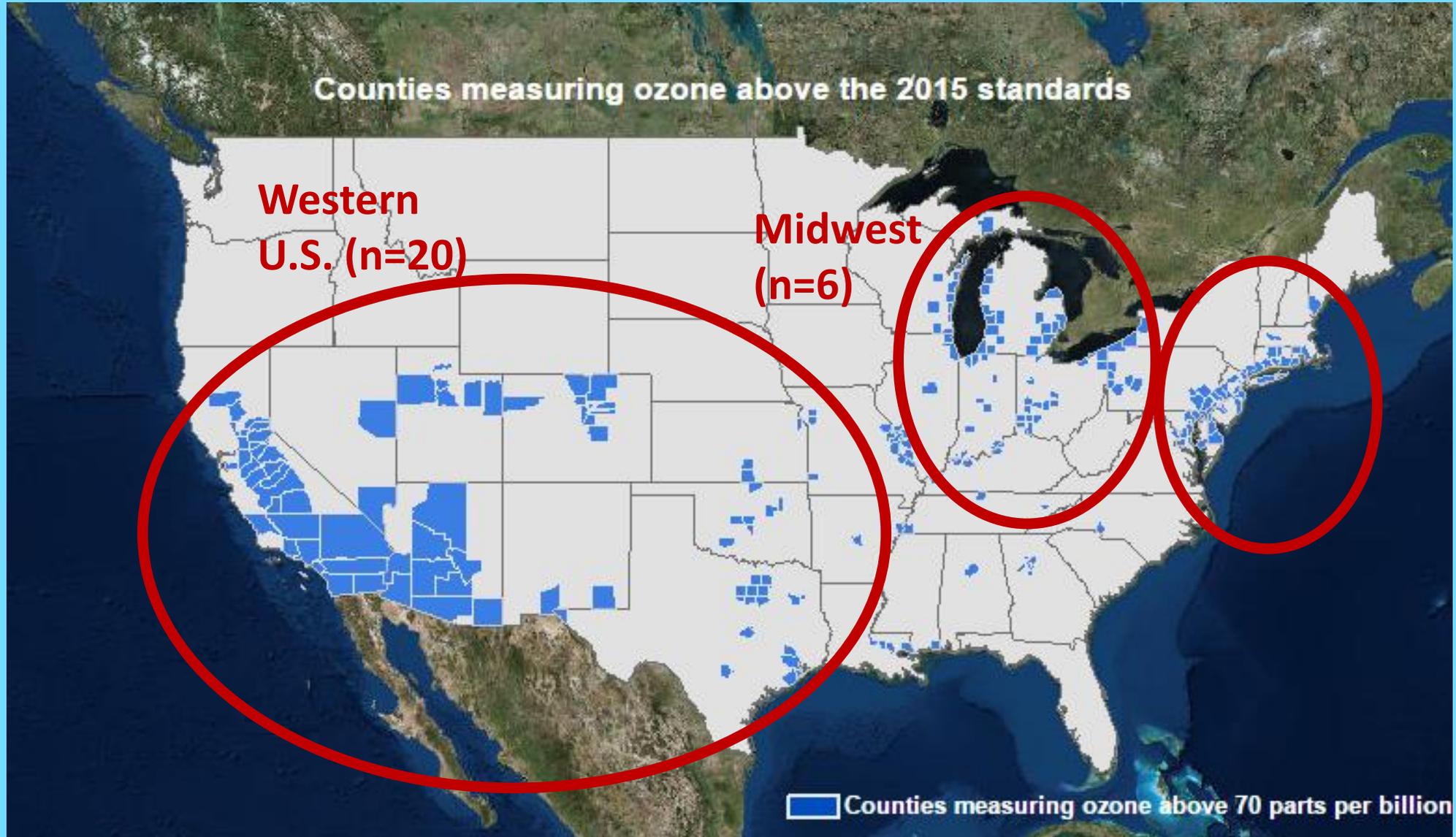


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~130 million people are exposed to unhealthy levels of O₃



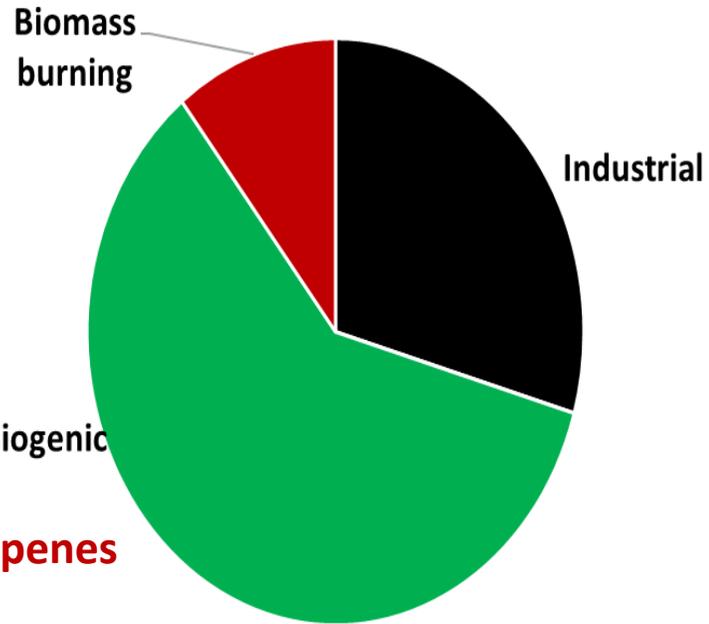
>130 million people are exposed to unhealthy levels of O₃



U.S. Emissions of NO_x and VOCs

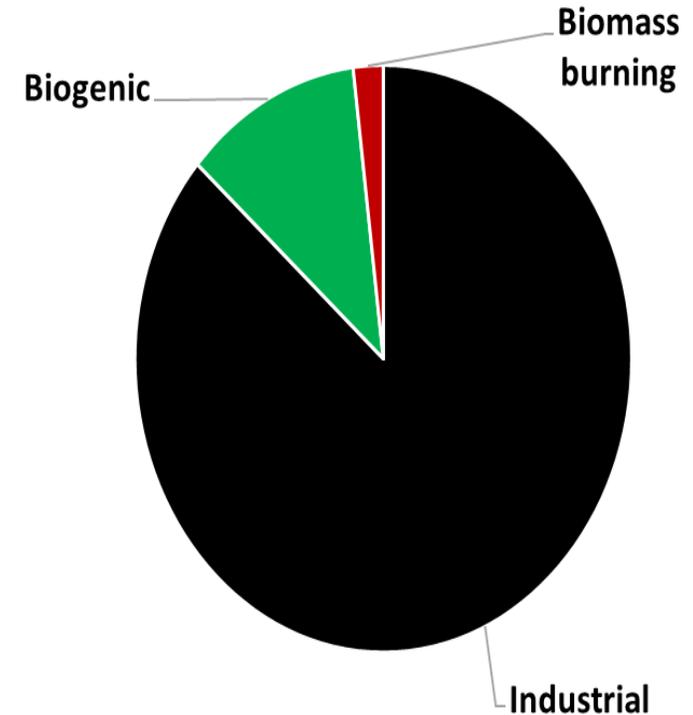
- NO_x is 87% from industrial/sources sources (30% from on-road vehicles)
- VOCs are 29% from industrial/anthro sources

U.S. VOC emissions



- Isoprene
- Mono-terpenes
- α -pinene
- etc

U.S. NO_x emissions

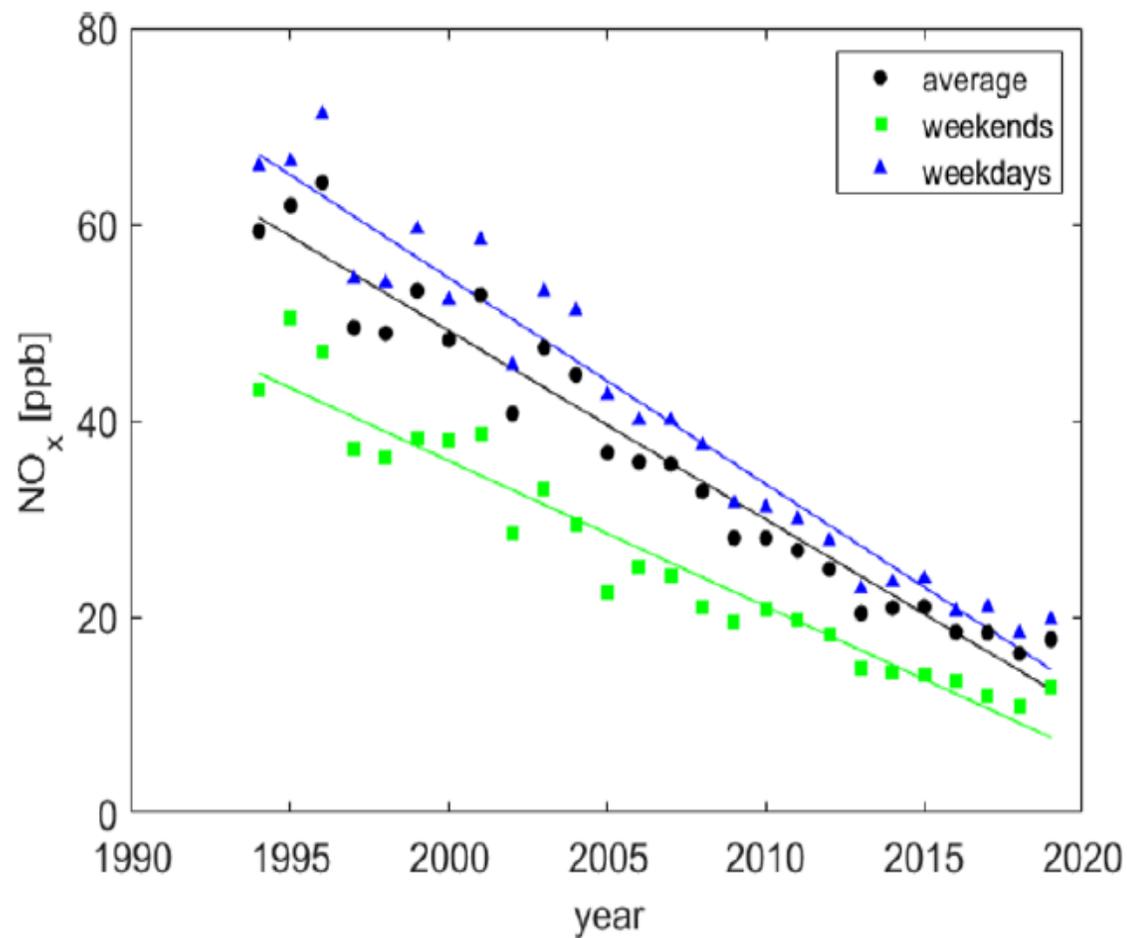


Source: EPA 2017 National Emission Inventory

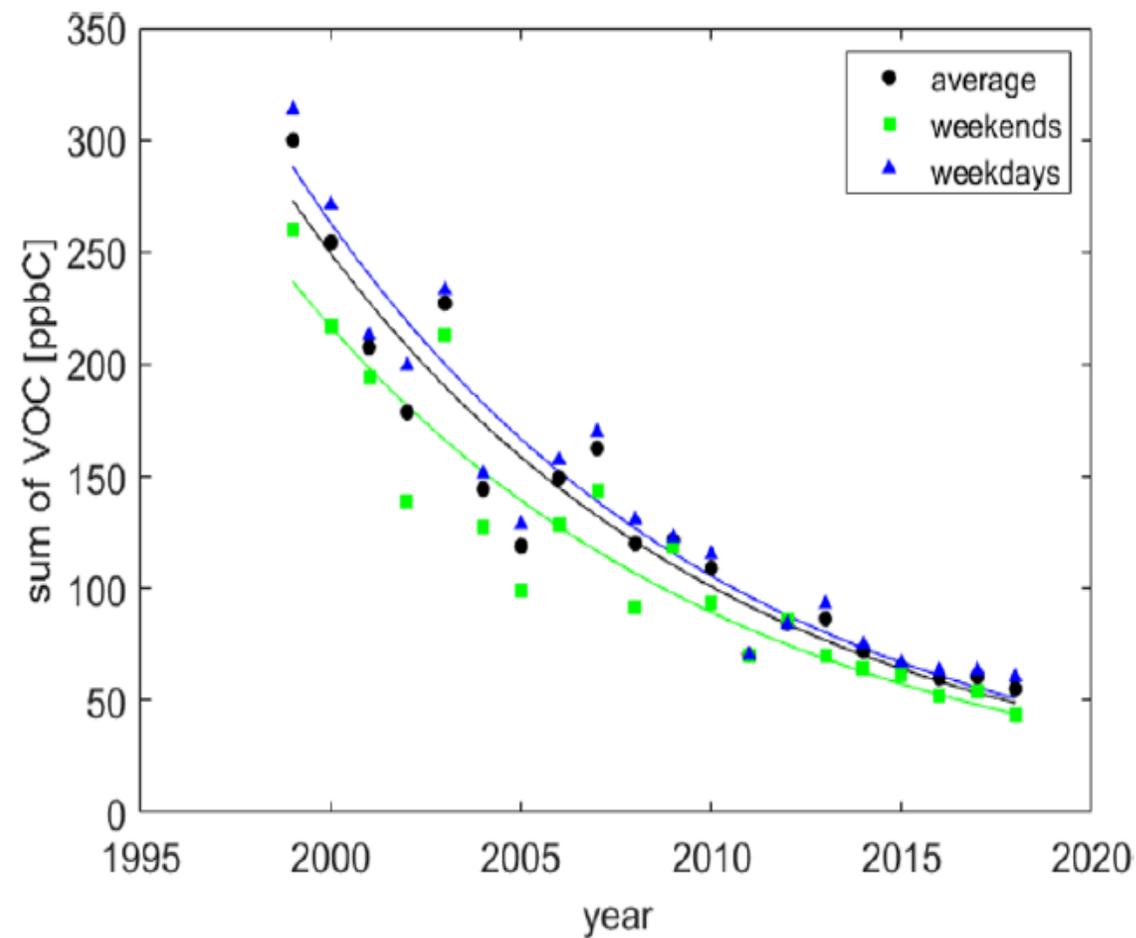


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NO_x and VOC emission trends, So.Cal

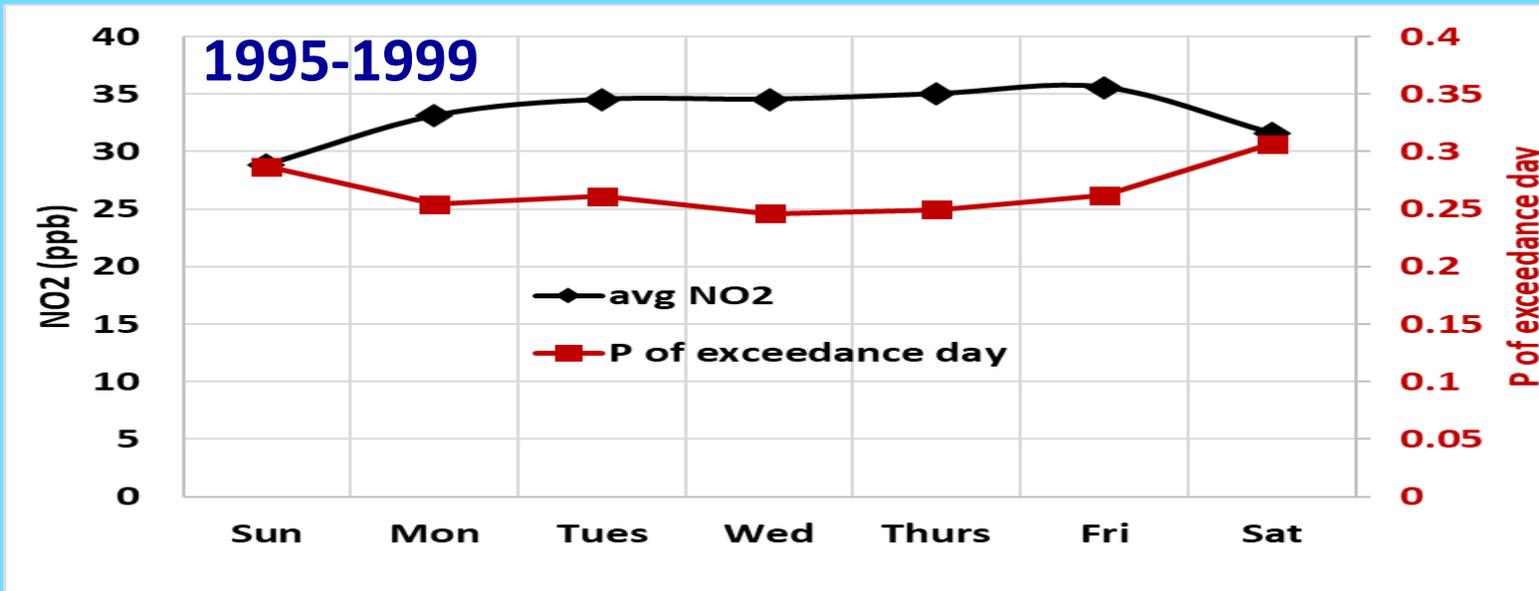


(a) NO_x



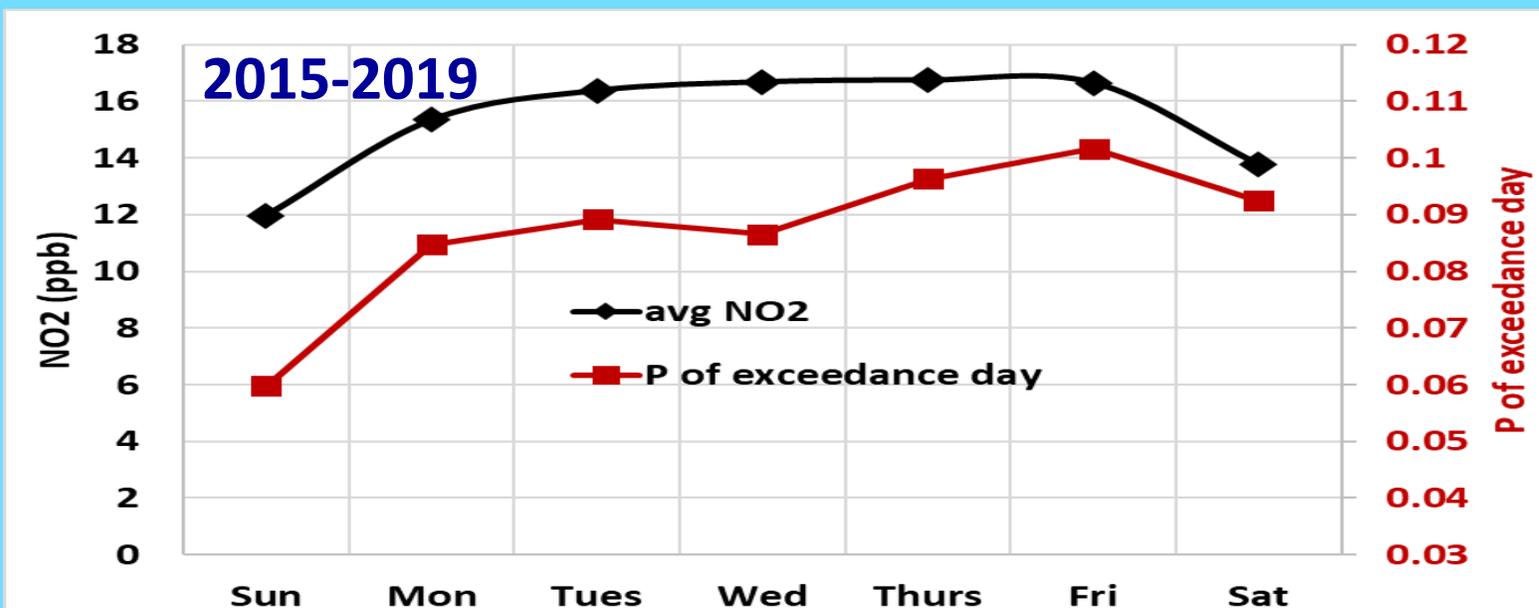
(b) VOC

NO₂ and probability of an O₃ exceedance by day of week, all sites



1995-1999, May-Sept:

NO_x saturation means that high O₃ days are more likely on weekends.



2015-2019, May-Sept:

NO_x limitation means that high O₃ days are more likely on weekdays.

This patterns holds in 30 out of 32 areas.
Only exception is LA/ Riverside area.



How did the Pandemic impact air quality?



Agenda No. 24

Emission Reductions and Air Quality Impacts from the COVID-19 Pandemic Response



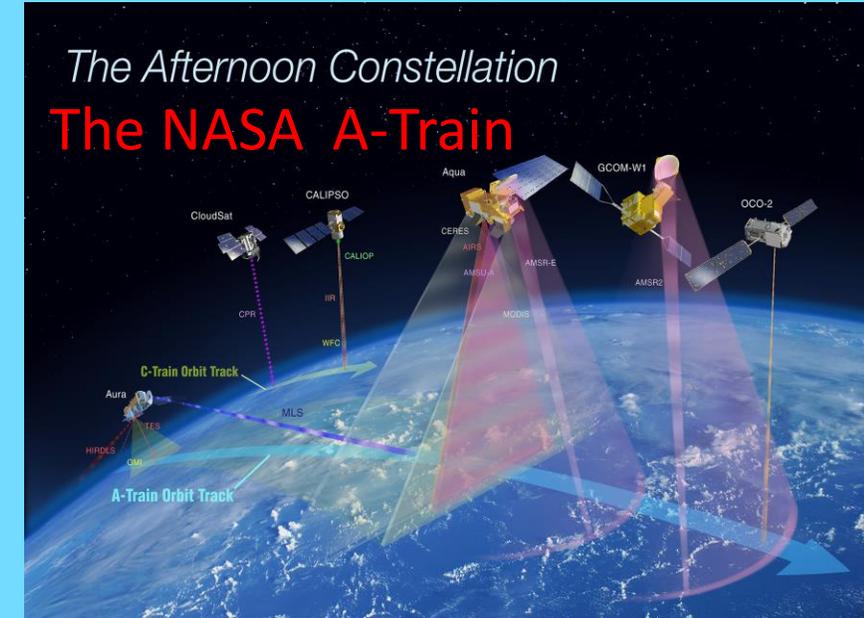
Governing Board Meeting
June 5, 2020



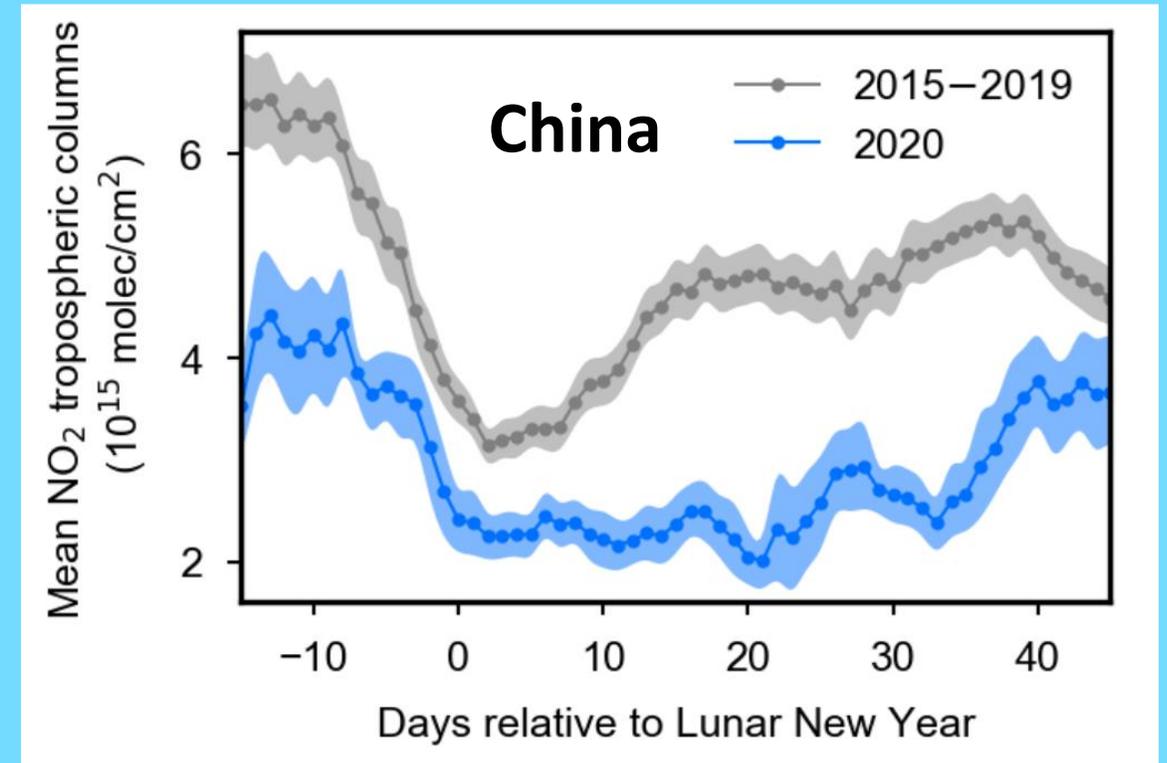
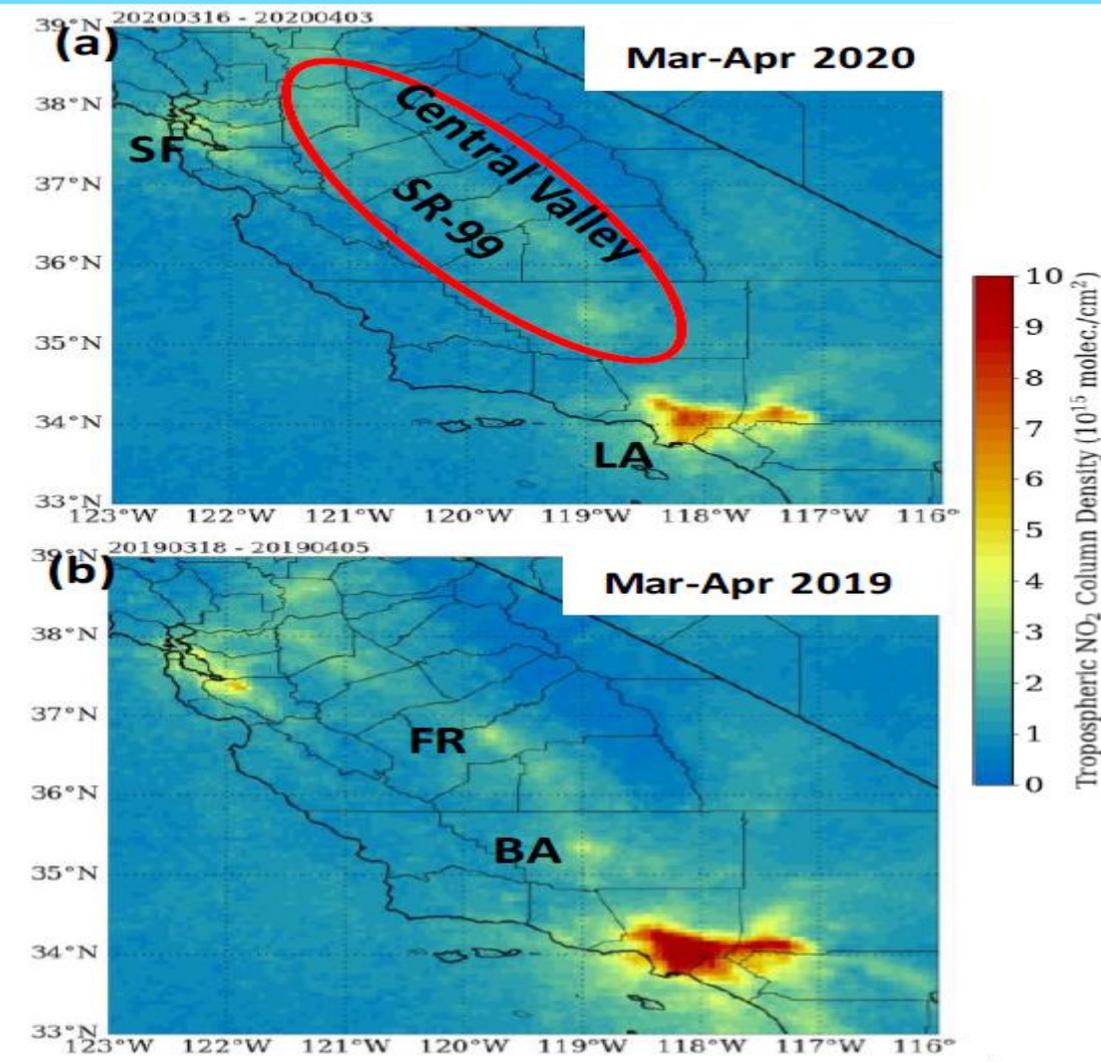
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Column NO₂ from OMI (Aura) or TROPOMI (Sentinel-5P)

- OMI and TROPOMI instruments for column NO₂ on polar orbiting satellites.
- Both use backscattered UV-Vis to measure NO₂.
- Need to subtract stratospheric NO₂ to get trop column.
- Daytime observations synched with equatorial crossing time.
- OMI = Launched in 2004 with res of 13 × 24 km² at nadir
- TROPOMI = Launched in 2017 with res of 7.2 × 3.5 km² at nadir



Analysis of satellite NO₂ data



Liu et al 2020

OMI data

Naeger and Murphy 2020
TROPOMI data



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Given the reductions in NO_2 can we use this to understand changes in O_3 ?

- Many papers show NO_2 reductions due to Covid lockdowns. But comments in some papers made little sense, like looking for O_3 reductions in March....
- Need to understand, which is limiting reagent NO_x or VOCs.
- Need to look for policy relevant changes in summer, when O_3 is elevated.

My strategy:

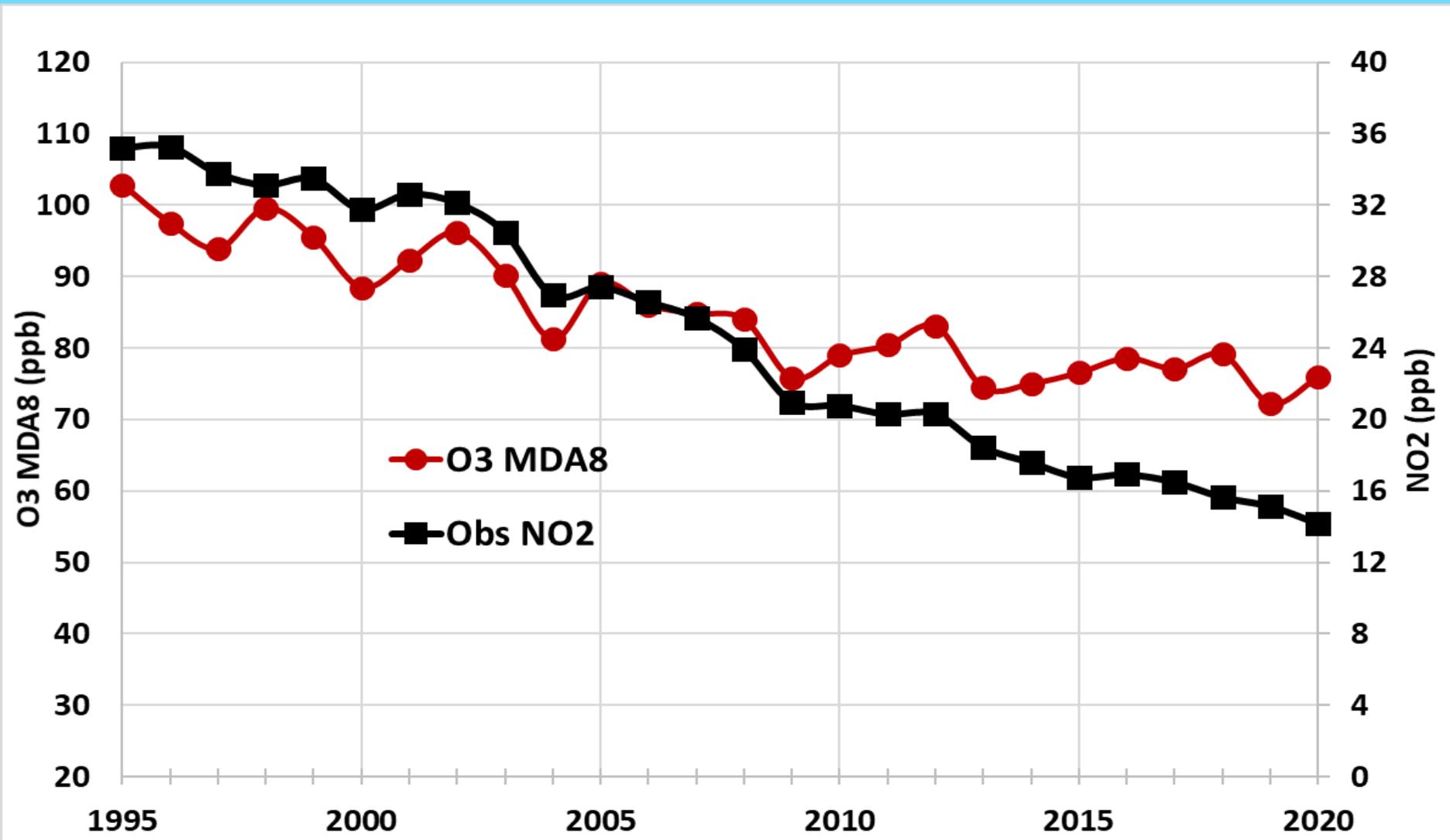
- Focus on 32 U.S. metropolitan regions that do not meet the O_3 standard (NAA: non-attainment areas)
- Look at annual fourth highest MDA8 at highest site by MSA (policy relevant metric).
- Use May-Sept mean of daily 1-hour maximum observed surface NO_2 to reflect emissions.
- Note that 1-hour NO_2 max is usually in early morning so these concentrations don't tell us much about the daytime concentrations. And because different regions have different # monitors, hard to compare actual concentrations between regions.

Sites (n=32)

Eastern U.S. (n=6)	Midwest (n=6)	West (n=20)	
Atlanta, GA	Chicago, IL-IN-WI	Butte Cty/Chico, CA	Phoenix-Mesa, AZ
Baltimore, MD	Cincinnati, OH-KY	Dallas-Fort Worth, TX	Riverside County, CA
Hartford CT CBSA.	Cleveland, OH	Denver, CO	Sacramento Metro, CA
New York, NY-NJ-CT	Detroit, MI	Dona Ana County, NM	San Antonio, TX
Phil., PA-NJ-DE-MD	Milwaukee, WI	Houston, TX	San Diego County, CA
Wash., DC-MD-VA	St. Louis, MO-IL	Imperial County, CA	San Francisco, CA
		Bakersfield, CA	San Joaquin Valley, CA
		Las Vegas, NV	San Luis Obispo, CA
		Los Angeles- CBSA	Sutter Butte/Yuba City, CA
		SLC, UT	Ventura County, CA



Fourth highest O₃ MDA8 and May-Sept mean NO₂

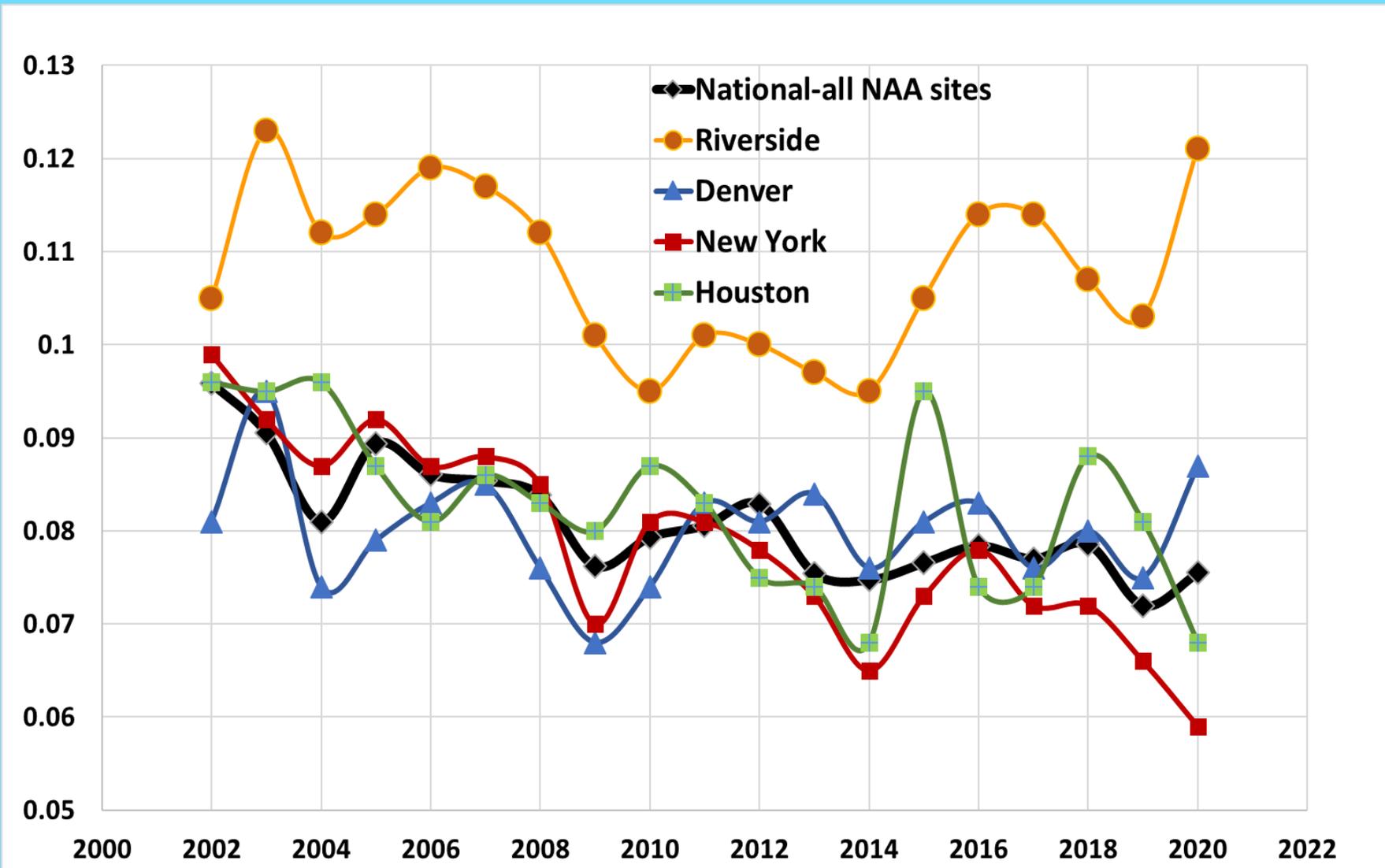


Averaged across the 32 U.S. NAAs, the policy relevant O₃ metric went up, not down!

What went wrong?



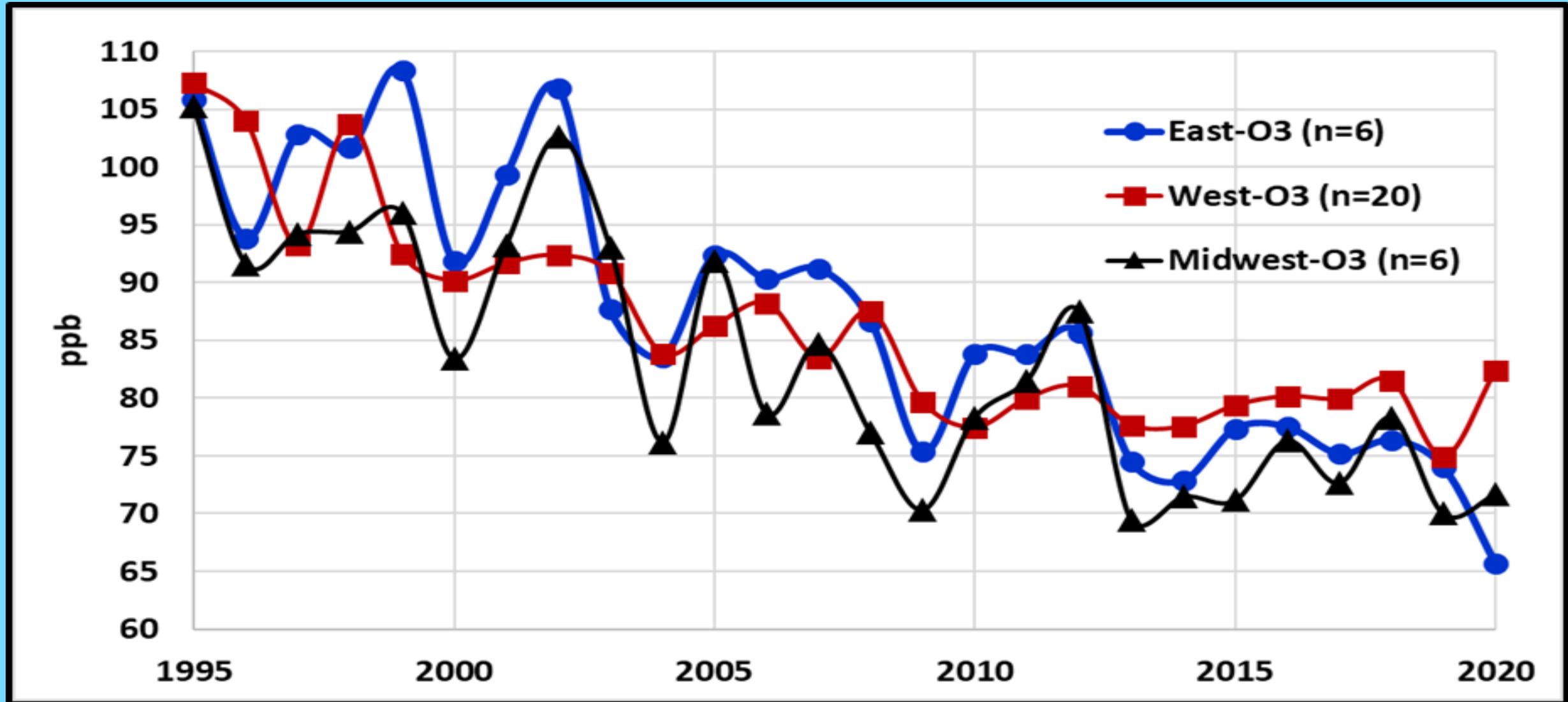
Annual 4th highest O₃ MDA8



Averaged over all all U.S. NAA, O₃ went up in 2020! Why are some sites up and some down?

- Hyp1: In NO_x limited case, reductions in NO_x will decrease O₃.
- Hyp2: Climate/temp variations.
- Hyp3: Large 2020 fires may have exacerbated O₃ at some sites.

Fourth highest O₃ MDA8 by region for U.S. non-attainment regions



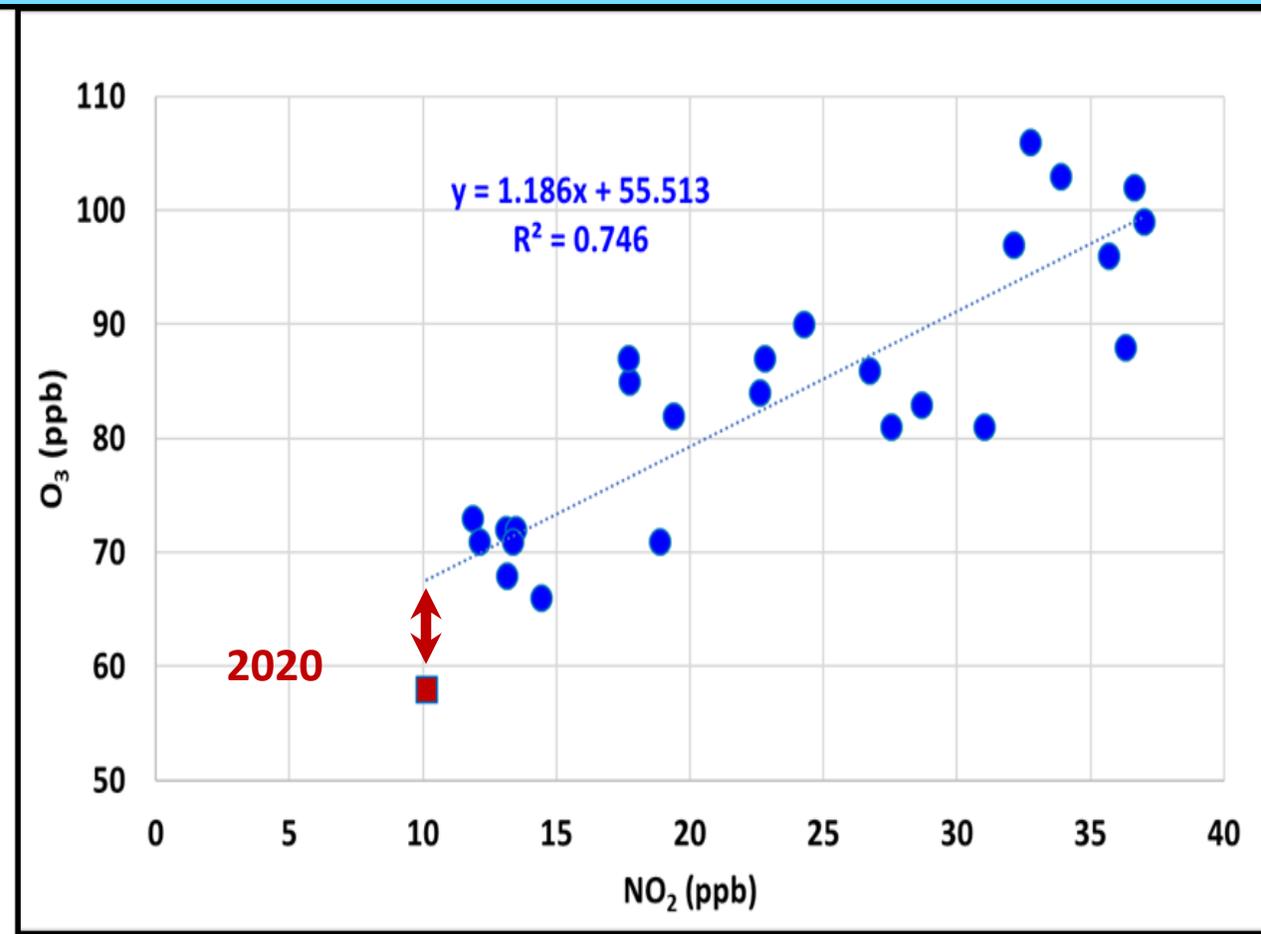
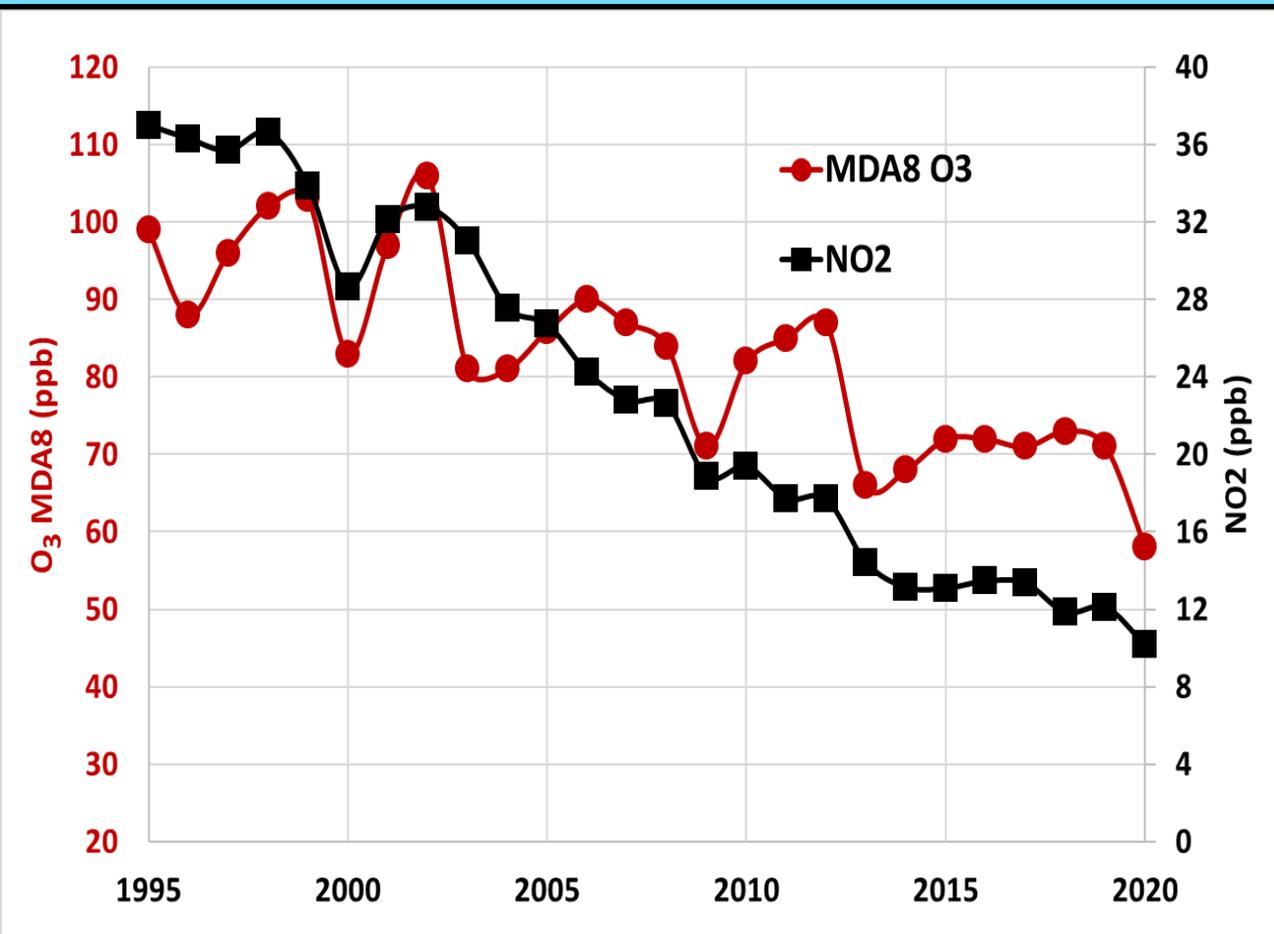
Twenty of the O₃ NAAs are in west, 12 in California. So this drives the national trend I showed a few slides back.

2020 and 2019 NO₂ in Wash D.C. metro region



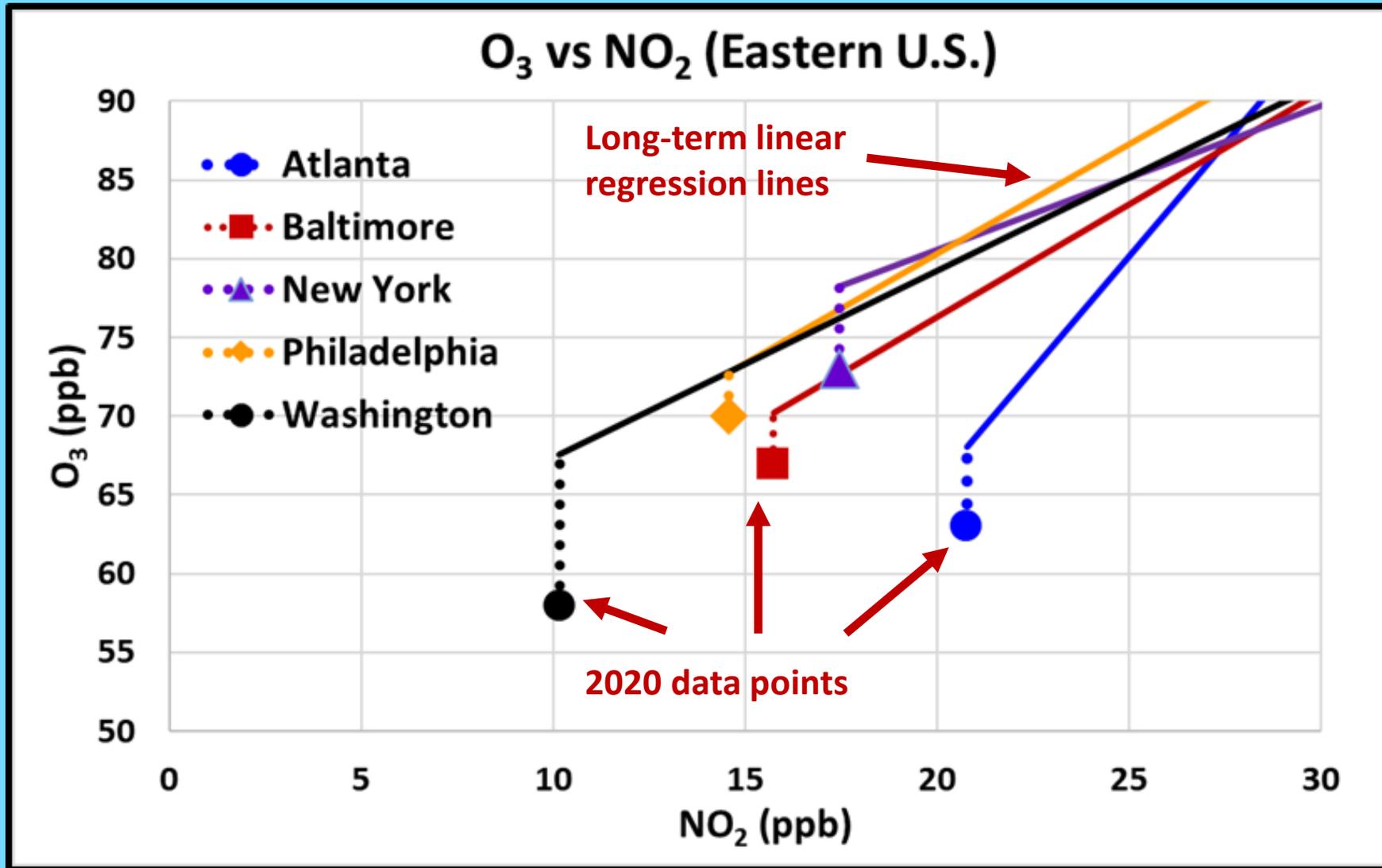
- Data from 5 surface NO₂ monitors in WDC region.
- NO₂ in 2020 was 28% lower in March compared to 2019; 16% lower for May-Sept.
- We found little connection between surface NO₂ changes and satellite NO₂ column change (see

Wash DC Fourth highest MDA8 O₃ and NO₂, 1995-2020



- 16% NO_x reduction in 2020 (compared to 2019) in May-Sept was consistent with a sig reduction in O₃.
- 2020 reductions were “super-linear”. What did other eastern cities show?

O₃ reductions in 2020 are “super-linear” at 5 Eastern cities



- No evidence for unusual met in summer 2020. Only explanation for the non-linear reduction is chemical.
- Note that except for NYC, all cities were at or below 70 ppb, most for the first time.

Ozone Production Efficiency (OPE)

- OPE is the number of O_3 molecules produced for each NO_x molecule converted to NO_y .
- Several researchers have reported that OPE is higher at lower NO_x concentrations (Kleinman 2002; Henneman et al 2017).



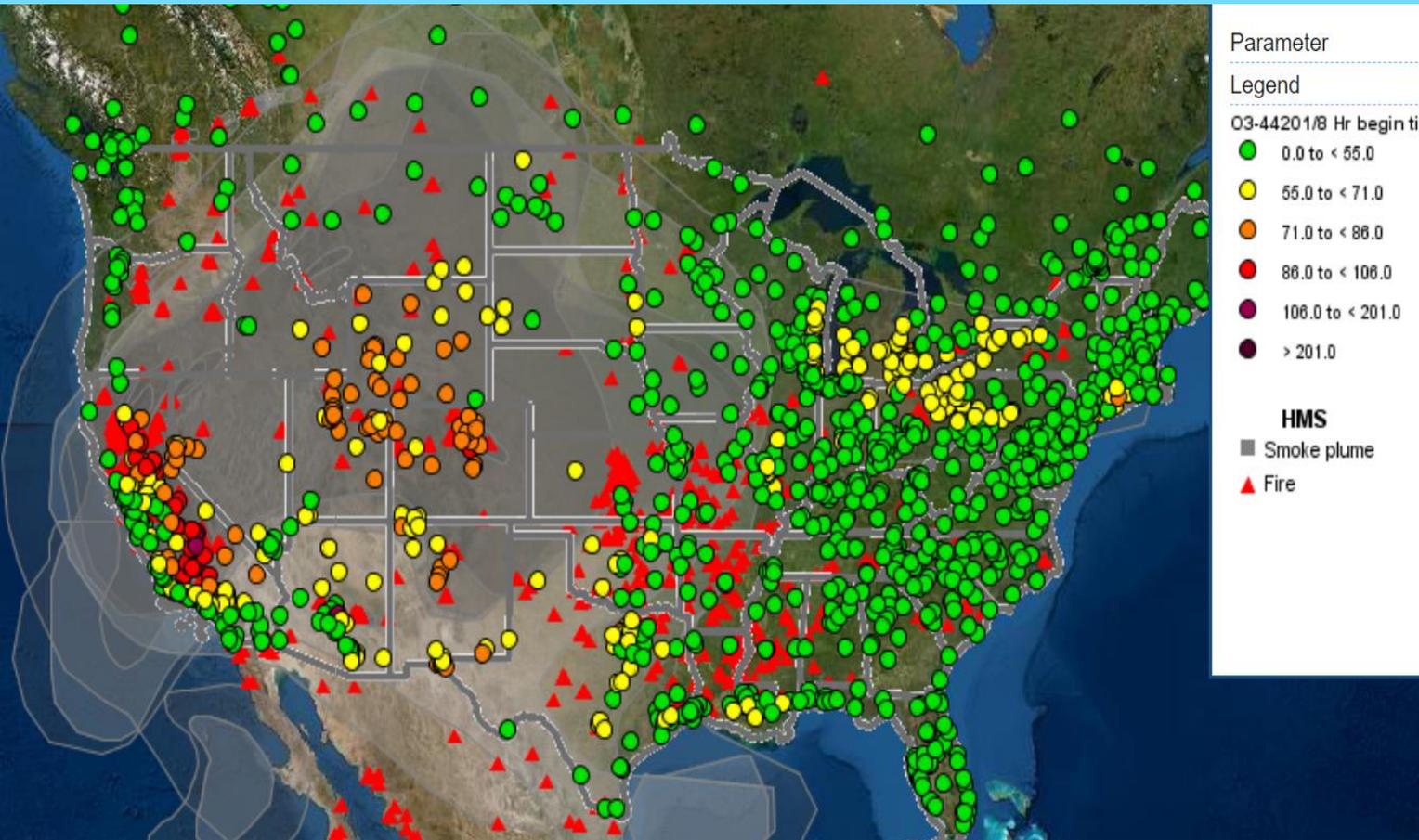
Henneman, *et al.*, Responses in ozone and its production efficiency attributable to recent and future emissions changes in the Eastern United States. *Environ. Sci. Technol.* 51, 13797–13805 (2017).

May-Sept. 2020 changes in NO_x and O₃ compared to 2019: Western U.S. (partial list)

Region	Surface NO ₂ (%)	Satellite NO ₂ (%)	O ₃ (ppb)
Denver CO	-4	-11	12
Dallas-Fort Worth, TX	-3	-2	6
Las Vegas, NV	-11	-5	10
Los Angeles CA	1	-17	15
Phoenix-Mesa, AZ	0	-6	9
Riverside CA	8	-10	18
Sacramento Metro, CA	-8	+13	15

**For the fourth highest MDA8 in the 20 areas in the W. US:
3 improved from 2019 to 2020 by an average of -8 ppb;
3 show no change from 2019 to 2020;
14 increased from 2019 to 2020 by an average of +11 ppb.**

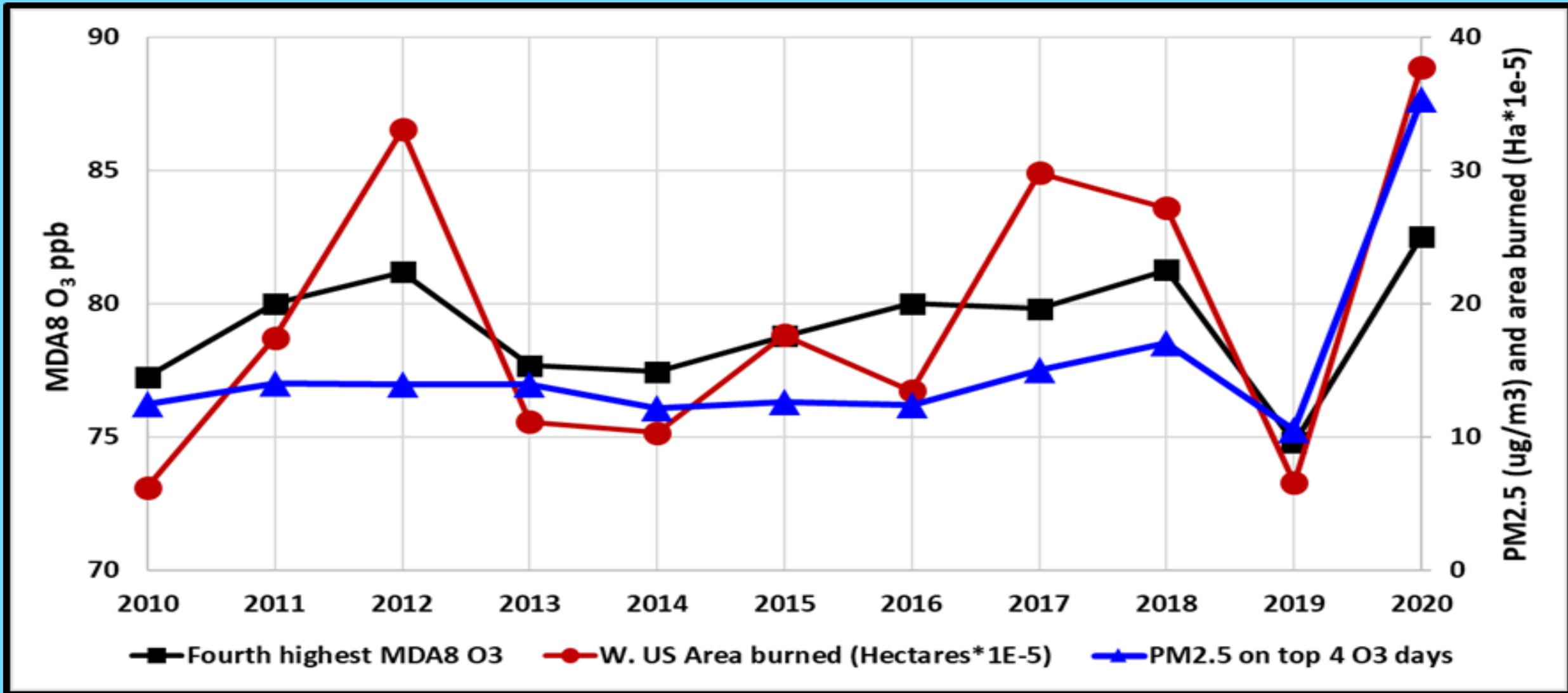
Photochemistry, O₃ and smoke.



MDA8 O₃ values for Aug 22, 2020

- We know that that smoke contributes to O₃ exceedances in urban areas, but processes poorly understood. (Jaffe et al 2004, 2008; 2018; 2020; Buysse et al 2019).
- Current work:
- Use machine learning/statistical models to tease out impacts of smoke (Gong et al 2017; McClure and Jaffe 2018; Jaffe and Gao 2021).
- Use photochemical box modeling (FOAM) to examine chemical processes in urban air+smoke (Ninneman et al 2021).

Fourth highest O₃ MDA8, PM_{2.5} and western area burned



Summary: How did 2020 Pandemic, fires and temps influence policy relevant O₃ in the U.S.?

- H1: In NO_x limited case, reductions in NO_x will decrease O₃: **True for Eastern U.S. sites.**
- H2: Climate/temp variations can offset NO₂ reductions: **Maybe true for Midwestern sites.**
- H3: Large 2020 fires may have exacerbated O₃ at some sites: O₃: **True for Western U.S. sites.**
- **Eastern U.S. sites showed significant reductions in both NO₂ and O₃.** The reductions were super-linear, due to an increasing OPE at lower NO_x concentrations. The 2020 NO_x reductions allowed most eastern U.S. NAAs to achieve a level that is consistent with meeting the O₃ NAAQS for the first time.
- **Midwest sites were more variable.** NO₂ dropped slightly in the O₃ season, but a hot summer may have caused O₃ to increase at most sites.
- **Western sites had a big increase due to fires.** 2020 had the largest area burned for the last 50 years. This caused massive air quality issues around the western U.S. and increased the fourth highest MDA8 O₃ values by 5-15 ppb. The increase in O₃ on smoke days is roughly 50/50 due to the increased temp and smoke photochemistry.
- **Key uncertainty: At this point we don't know how much VOCs might have changed in 2020 or whether they still partially control urban photochemistry. Evidence suggests we've mostly moved into NO_x control, maybe... Stay tuned for more on this.**

Spares

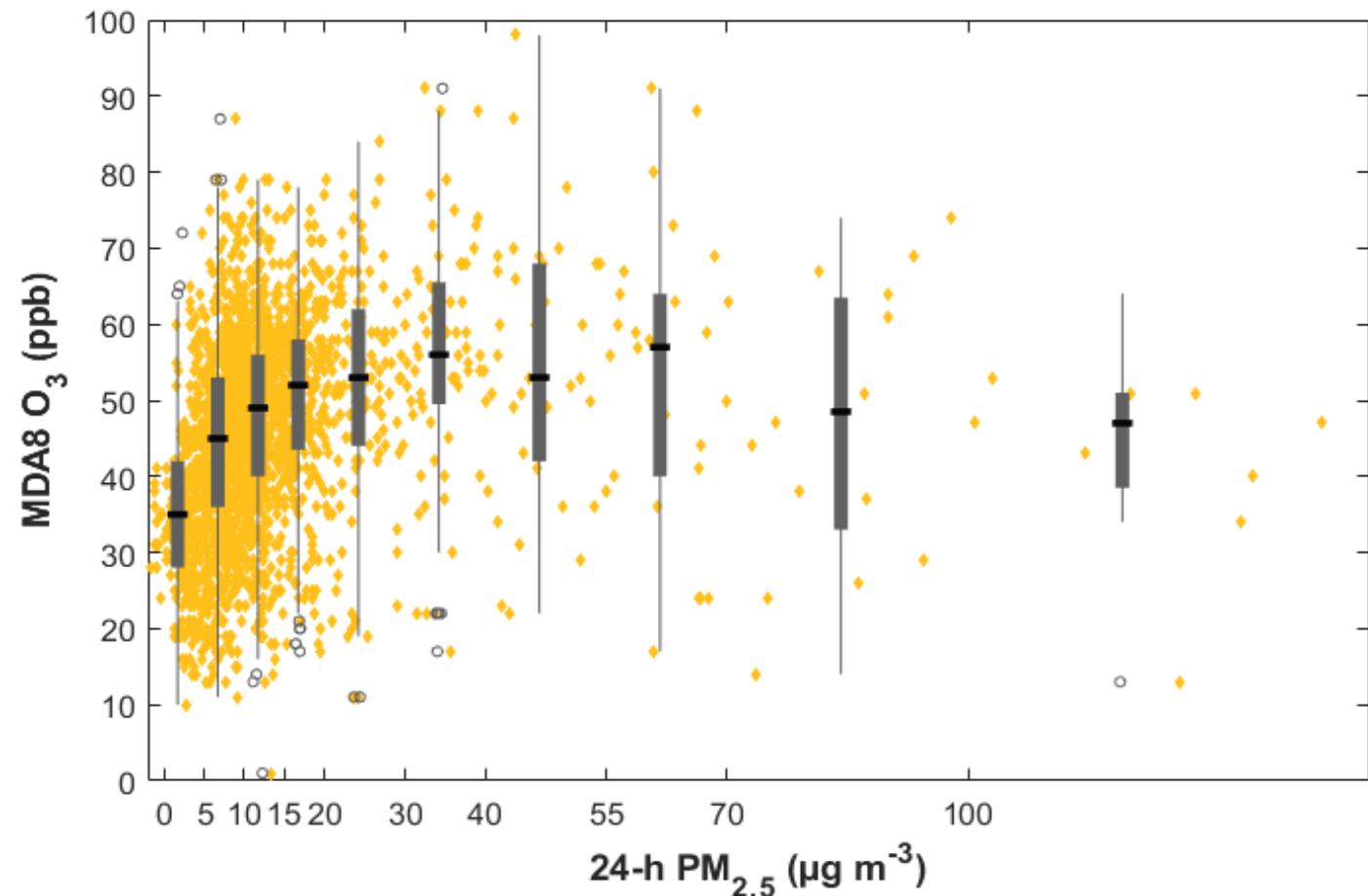


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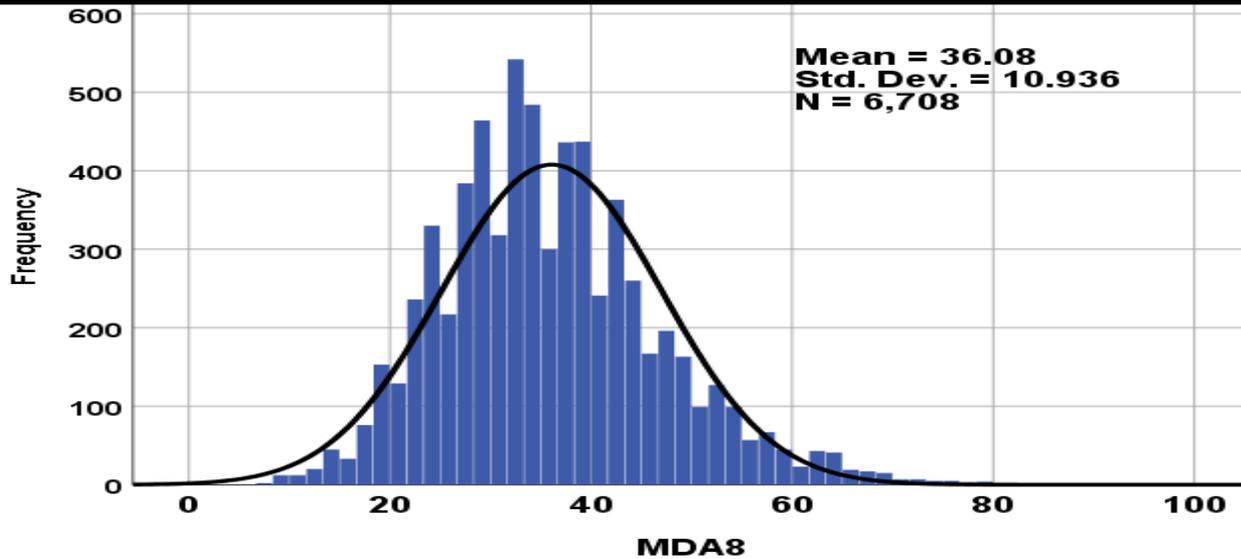
Very high PM_{2.5} does not lead to high O₃ in smoke



- Role of transport time and PM levels (McClure et al 2018; Buysse et al 2019); Role of UV Photolysis (Baylon et al 2018; Alvarado et al 2015).
- Role of Temperature (Gao et al 2020).
- Use of machine learning to predict O₃ MDA8 in smoke (Gong et al 2017; Gao et al 2020)
- Why not high O₃ at high PM_{2.5}?
Heterogenous chemistry?
Reduction in photolysis?

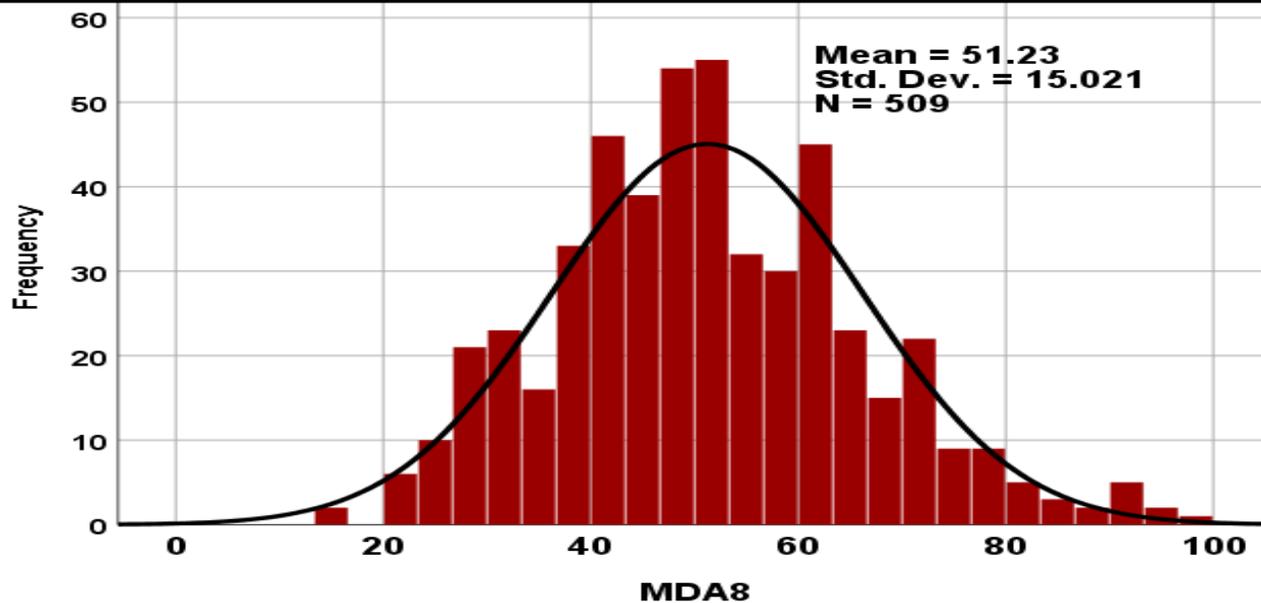
Buysse et al 2019.

Max daily 8-hour O₃ in and out of smoke for 5 cities in the PNW: May-Sept. 2007-2017.



No smoke days:

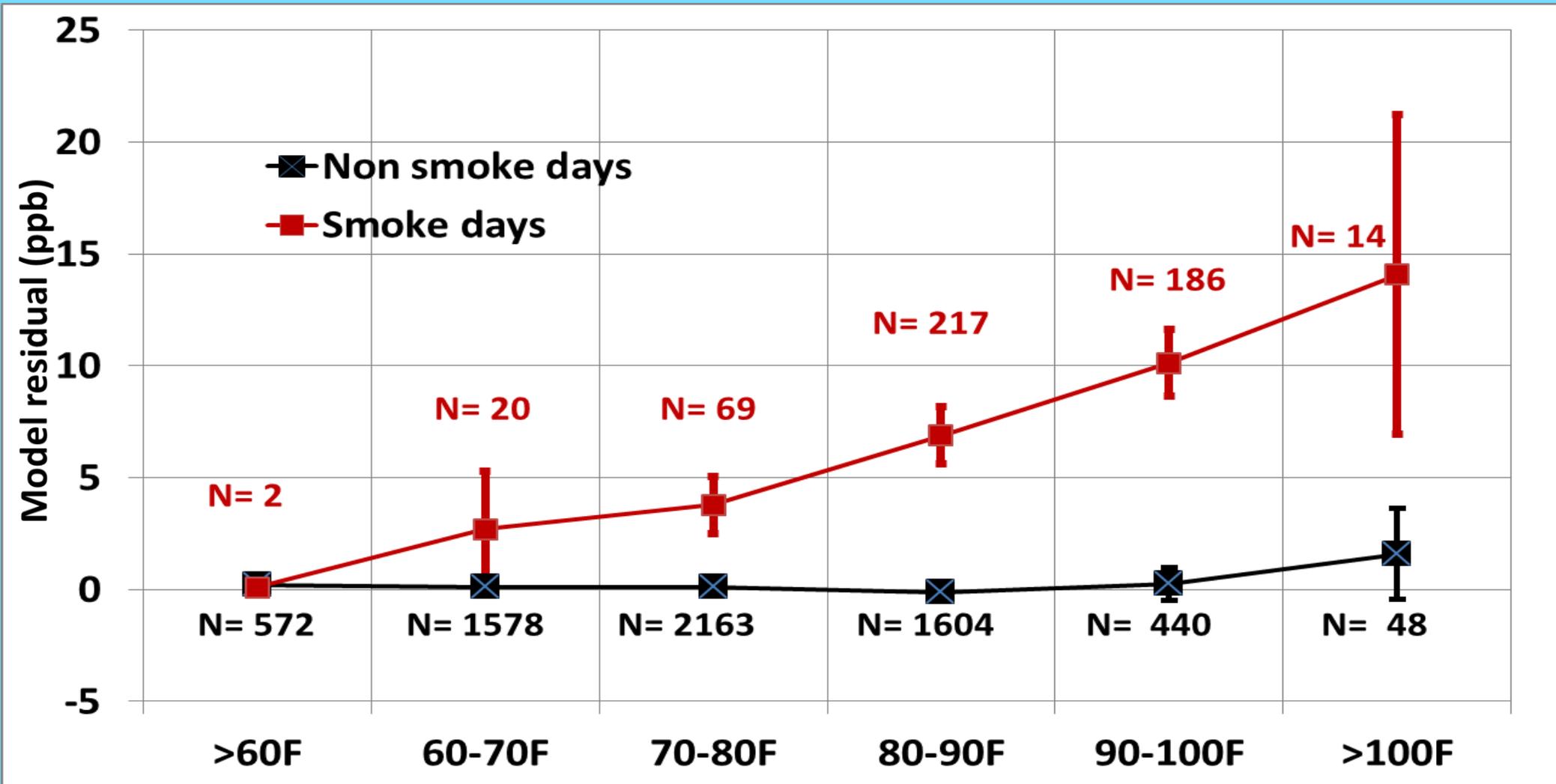
- Average MDA8 = 36 ppb
- Fraction of days that have MDA8 g.t. 70 ppb = 0.4%



Smoke days:

- Average MDA8 = 51 ppb.
- Fraction of days that have MDA8 g.t. 70 ppb = 10.1%.
- Smoke days are 40% of all days with MDA8 > 70
- Using a machine learning algorithm, we find that smoke contributes about 8 ppb, on average, to the MDA8, but with lots of variability (Jaffe et al 2021).

MDA8 residuals vs daily max temp (mean +/- CI, ppb)



This demonstrates that the GAMs correctly predict O_3 as a function of temperature in the absence of smoke.

Jaffe et al 2021



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MDA8 O₃ temp relationship, SLC-BV site, May-Sept 2006-2018

